ORIGINAL PAPER

Isolation of the opdE gene that encodes for a new hydrolase of Enterobacter sp. capable of degrading organophosphorus pesticides

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Abstract Microbial enzymes that can hydrolyze organophosphorus compounds have been isolated, identified and characterized from different microbial species in order to use them in biodegradation of organophosphorus compounds. We isolated a bacterial strain Cons002 from an agricultural soil bacterial consortium, which can hydrolyze methyl-parathion (MP) and other organophosphate pesticides. HPLC analysis showed that strain Cons002 is capable of degrading pesticides MP, parathion and phorate. Pulsed-field gel electrophoresis and 16S rRNA amplification were performed for strain characterization and identification, respectively, showing that the strain Cons002 is related to the genus *Enterobacter* sp.

which has a single chromosome of 4.6 Mb and has no plasmids. Genomic library was constructed from DNA of *Enterobacter* sp. Cons002. A gene called *opdE* (Organophosphate Degradation from Enterobacter) consists of 753 bp and encodes a protein of 25 kDa, which was isolated using activity methods. This gene *opdE* had no similarity to any genes reported to degrade organophosphates. When kanamycin-resistance cassette was placed in the gene *opdE*, hydrolase activity was suppressed and *Enterobacter* sp. Cons002 had no growth with MP as a nutrients source.

Keywords Methyl-parathion \cdot Organophosphorus pesticides \cdot *Enterobacter* sp. Cons $002 \cdot opd$ E Gene \cdot Degradation

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Introduction

Synthetic organophosphate (OP) compounds have been widely used as pesticides and have played an important role in raising agricultural productivity and controlling pests. However, these compounds are often toxic in non-target organisms, and their transformation products are frequently found in the environment (Van and Pletschke 2010).

These compounds irreversibly inhibit acetylcholinesterase (AchE), a key enzyme in the central nervous systems of insects and non-target organisms, including humans (Sogorb and Vilanova 2002; Zhang et al.



2006). Effective methods for degrading organophosphates are needed to ensure that the health of neither humans nor the environment will be compromised. Previous forms of disposal for OP compounds have consisted of chemical or physical treatments (Chen et al. 2000), but they produce toxic residues; therefore, other environmentally friendly technologies, like the use of microorganisms or their enzymes, are now being considering for eliminating organophosphorus residues.

Several enzymes called phosphotriesterases (PTE) isolated from different microorganisms have been found to hydrolyze and detoxify OPs. This reduces OP toxicity by decreasing the ability of OPs to inactivate AchE (Ghanem and Raushel 2005; Singh and Walker 2006; Porzio et al. 2007; Theriot and Grunden 2010; Shen et al. 2010). These enzymes specifically hydrolyze phosphoester bonds, such as P–O, P–F, P–NC, and P–S, and the hydrolysis mechanism involves a water molecule at the phosphorus center (Ortiz-Hernández et al. 2003). Different microbial enzymes with the capacity to hydrolyze MP have been identified, such as organophosphorus hydrolase (OPH; encoded by the opd gene), methyl-parathion hydrolase (MPH; encoded by the mpd gene), and hydrolysis of coroxon (HOCA; encoded by the hocA gene), which were isolated from Flavobacterium sp. (Sethunathan and Yoshida 1973), Plesimonas sp. strain M6 (Cui et al. 2001) and Pseudomonas moteilli (Horne et al. 2002), respectively.

We have previously described the characterization of a bacterial consortium that had been isolated from agricultural soils with a history of pesticide application (Yáñez-Ocampo et al. 2009). The isolated consortium was adapted for several weeks to a mineral medium without an additional carbon source. The mineral medium contained untreated OP pesticide wastes generated from livestock-dipping operations that used an OP pesticide mixture. The consortium was subcultured and continuously enriched throughout the application of selective pressure. The consortium consisted of 14 bacterial strains, 12 of which were capable of degrading MP. Out of the 12, strain Cons002 showed the greatest ability to hydrolyze MP and was selected for further research (Specific whole cell activity = 2.03 ± 0.18 mU/mg protein).

In the present study, the strain Cons002 was characterized. The *opdE* gene was isolated, sequenced, and effectively cloned into *Escherichia coli*. We found that the *opdE* gene is responsible for

the degradation of organophosphorus pesticides. A phylogenetic analysis permitted the study of the evolution of the opdE gene.

Materials and methods

Reagents

MP (*O*,*O*-dimethyl *O*-4 nitrophenyl phosphorothioate), parathion (*O*,*O*-diethyl-*O*-*p*-nitrophenyl phosphorothioate), phorate (*O*,*O*-diethyl *S*-[(ethylthio) methyl] phosphorodithioate), cadusafos (*S*,*S*-di-*sec*-butyl *O*-ethyl phosphorothioate) and butiphos (*S*,*S*,*S*-tributyl phosphorotrithioate) were purchased from Chemservice (99% purity). All of the other chemicals were of reagent grade and were obtained from J. T. Baker, Mexico City. Restriction enzymes were acquired from New England Biolabs.

Bacterial strains

The bacterial strains used were E. coli XL1-Blue (recA1 endA1 gyrA96 thi-1 hsdR17 $[r_{\kappa}^{-} m_{\kappa}^{+}]$ supE44 relA1 lac [F' proAB lacI^qZ∆M15 Tn10 Tet^r]); E. coli DH5\alphasupE44 DlacU169 (f80lacZ DM15) hsdR17 recA1 endA1 gyrA96 thi-1 relA1 (Sambrook and Russell 2001); strain Cons002 (Wild-type strain, MP^r, Ap^r) (Yáñez-Ocampo et al. 2009); Hydro1257 (DH5pKS + opdE), Enterobacter opdE- (Enterobacter opdE::Km^r) and Enterobacter pKSopdEGM (Enterobacter opdE:: Km^r pKS + opdE) (this study); and BL21 (DE3) (hsdS gal [\lambdacIts857 ind1 Sam7 nin5 lacUV5-T7]). The plasmids pJET (sequencing vector Ap^r [Fermentas]), pBluescript II K^+ (pKS +) (oriR101, repA101 (ts); araBp-gam-bet-exo [Stratagene]), pKD78 (Red helper plasmid, Ts; Cm^r), pKD4 (oriRy; Km^r, Ap^r) (Datsenko and Wanner 2000) and pET-22b (+) (Novagen) were used.

Culture medium composition

For the cultivation of the Cons002 strain, minimal medium (MM) was modified from a recipe described by Brown (1980) and had the following composition (g/l): 2.92 KH₂PO₄, 2.74 K₂HPO₄, 0.20 MgSO₄· 7H₂O, 2.0 KNO₃, and 0.99 (NH₄)₂SO₄. It also contained glucose at concentration of 50 mg/l and 2 ml/l of a trace element solution composed of (g/l)



2.8 H_3BO_3 , 2.55 $MnSO_4 \cdot H_2O$, 0.20 $CuSO_4 \cdot _5H_2O$, 2.43 $CoCl_2 \cdot _6H_2O$, and 0.25 $ZnSO_4 \cdot _7H_2O$ at pH 7 \pm 0.05 and 30°C. The MP was dissolved in methanol and added to the media to a final concentration of 20 mg/l. *Escherichia coli* was grown at 37°C in LB medium with ampicillin, kanamycin, chloramphenicol and gentamicin at concentrations of 100, 60, 30 and 15 μ g/ml, respectively. When necessary, the Cons002 strain was grown in Tripticasein soy (TS) broth (Bioxon, Becton–Dickinson, Mexico).

Degradation of MP by Cons002

The Cons002 strain was cultured overnight in 3 ml of LB medium containing 100 μg/ml ampicillin. The culture was then used to inoculate 50 ml of MM containing 50 mg/l glucose and 20 mg/l MP with an initial OD₆₀₀ of 0.05. Samples (1 ml) were collected at 0, 4, 8, 12, 16, 20, 24 and 28 h after inoculation for viable count analysis (CFU/ml). Chemical analyses of residual MP and p-nitrophenol (PNP) concentrations were carried out by HPLC using the Waters 2487600 ultraviolet (UV) detector, a Dual Absorbance Detector, and an EC 250/4 Nucleosil 120-5 C18 HPLC column. The mobile phase was methanol/water (60/40, v/v), and a flow rate of 0.8 ml/min was used. MP and PNP were detected at 275 nm. Simultaneously, the glucose concentrations in the culture supernatants were measured using an YSI 2700 SELECT biochemistry analyzer (YSI Inc., Yellow Springs, OH) equipped with a glucose oxidase membrane (YSI 2365). To determine the concentration of glucose in each culture supernatant, a 1 ml sample was taken every 4 h. The sample was centrifuged for 1 min, and the supernatants were used to determine the concentration of glucose. The same medium without inoculant was used as a control.

Utilization of other OP pesticides as substrates

The Cons002 strain was cultured in MM with 50 mg/l glucose and 20 mg/l of the parathion, cadusafos, butiphos or phorate pesticides. The cultures were incubated under previously described conditions; non-inoculated media were also used as controls. After 28 h of incubation, the residual concentration of pesticides was analyzed by HPLC under the same conditions described above. Parathion was detected at

275 nm, while phorate, cadusafos and butiphos pesticides were determined at 230 nm.

Enzyme assay

Cells were grown in MM with 20 mg/l of MP for 48 h at 28°C and agitated at 100 rpm. After being harvested, the cells were suspended in 2 ml of distilled water and sonicated at 25% efficiency for 75 s (3 pulses of 25 s each) in a sonicator (Branson Sonifier 450). After sonication, the solution was centrifuged $(17,000 \times g$ for 10 min), and the intracellular fractions were tested for hydrolase activity. The protein concentration was determined, using bovine serum as standard, by the method developed by Lowry et al. (1951).

Hydrolysis of MP was measured spectrophotometrically by monitoring the production of PNP at 405 nm ($\varepsilon_{\rm max}=17,000~{\rm M}^{-1}~{\rm cm}^{-1}$) for 6 h at 28°C; assays were performed in 20 mM Tris–HCl (pH 9). Enzyme activity was defined as the amount of enzyme required to release 1 µmol of PNP per min at 37°C. The data are reported as specific activity (mU/mg of protein).

Bacterial identification

Chromosomal DNA was extracted by the method developed by Miller et al. (1988) and used for PCR amplification of the 16S rRNA gene. The 16S rRNA gene of the bacterial strain Cons002 was amplified universal primers 63Fw (5'-CAGwith GCCTAACACATGCAAGTC-3') and L1401 Rv (5'-CGGTGTGTACAAGACCC-3') (Lopez et al. 2003). The amplified products were purified, and the 1.4 kb fragments were ligated into the pJET plasmid. The recombinant plasmid was transformed into competent E. coli DH5 α cells. The inserted fragments were sequenced with the primers 63Fw, L1401 and U968 (internal primer; 5'-AACGCGAAGAACCTTAC-3') by the Sequence Unit of Biotechnology Institute UNAM, Mexico. The sequence data were analyzed using the BioEdit sequence analysis software (http:// www.mbio.ncsu.edu/bioedit/bioedit.html).

A modified protocol for pulsed-field gel electrophoresis (PFGE) (Smith and Cantor 1987) was used for bacterial identification. The sample plugs were placed in 10 ml of fresh lysis solution (10 mM Tris–HCl, pH 7.2, 50 mM NaCl, 100 mM EDTA, 0.2% deoxycholate, 0.5% sodium lauroyl sarcosine, 20 µg/ml RNase



(DNase-free) and 1 mg/ml lysozyme) to disrupt the cells. The DNA fragments were separated by electrophoresis on a 0.6% agarose gel (Bio-Rad) using a 0.9× TAE buffer (20 mM NaCl, 50 mM EDTA, and 10 mM Tris–HCl, pH 7.2) in a contour-clamped homogeneous electric field device (CHEF-DR II) from Bio-Rad; the pulse time was increased from 2,600 s to 85 h at 1.8 V, 14°C and 6 V/cm. The electrophoresis was performed using alternating pulses at a 106° angle, with a 5–20 s pulse time gradient over a period of 85 h.

Isolation and sequence of the opdE gene

A gene library was constructed from genomic DNA. A partial digestion was performed with the enzyme Sau3AI to obtain randomly digested chromosomal DNA. The 1–5 kb fragments were recovered using a cleanup kit (High Pure PCR-Roche) and ligated into a pKS⁺ vector that had been digested with BamHI and treated with calf intestinal alkaline phosphatase (Roche). The ligation products were transformed into E. coli XL1-Blue competent cells, and the bacteria were plated on LB agar plates containing 100 μg/ml ampicillin for selection. The library was screened for clones harboring hydrolase activity towards MP. Positive clones have been sequenced in both directions with the primers T7 (5'-AATACGACTCACTATAG-3') and T3 (5'-ATTAACCCTCACTAAAG-3') (Stratagene) and the internal primers opdEFw (5'-TTTACCCGCT CACGCGGTCGTCC-3') and opdERV (5'-GCTCGA GATCATTGCCGCACCAGAATAC-3') to ensure the absence of sequencing errors.

Construction of a defective opdE mutant

A deletion of the *opdE* gene was made in the Cons002 strain using Red-mediated recombination (Datsenko and Wanner 2000). The *opdE* deletion was constructed using the kanamycin resistance *cassette* generated by PCR from the pKD4 template using primers opdEH1P1 (5'-TCACCCAGGAAGGTGTTTATGAGCAAAAAA ACGCAGCATTTCTGTAGGCTGGAGCTGCTTCG-3') and opdEH2P2 (5'-TCGCCATGAACTTTTCATTAAG TGGATCTCCGCGCTGAGTGGCATATGAATATC CTCCTTAG-3'). The strain mutant was selected for kanamycin resistance, and its specific activity was

determined. The mutant was confirmed by PCR using genomic DNA from the mutant strain as a template. The primers opdEH1P1 and opdEH2P2 amplified the predicted 1,500 bp fragment, while internal primers for the *opdE* gene, XHyFw (5'-TGCTCGAGATGAGCA AAAAAACGCAGCATTTCTCG-3') and BHyRv (5'-AGGATCCTCATCATAAGTGGATCTCGGC GC-3'), could not be used to amplify products (data not shown).

Nucleotide sequence accession number

The nucleotide sequence of the *opdE* gene was deposited into the DDBJ database with accession number JF837335.

Construction and plasmid complementation of the *Enterobacter* strain lacking the *opdE* gene

To construct the plasmid pKSopdE, the opdE gene was amplified using PCR. The upstream primer, opdERBS (5'-AACTCGAGAAGAAGGAGATATCAT**ATG**AG CAAAAAACG-3'), contained an XhoI restriction site (underlined bases) and a strong RBS (in italic) at the opdE start codon (in bold), while the downstream primer, opdERV (5'-AAGCTTCATCATAAGTGG ATCTCCGCGCT-3'), contained a HindIII restriction site (underline bases) at the stop codon (in bold). The 0.783 kb fragments were recovered using a clean-up kit (High Pure PCR-Roche) and ligated into pKS⁺ digested with XhoI and HindIII. The ligation products were transformed into E. coli DH5 α competent cells, and the bacteria were plated on LB agar plates containing 100 µg/ml ampicillin. The gentamicinresistance cassette was obtained from the pBSL142 plasmid (Alexeyev et al. 1995) using the XbaI restriction site, and the resulting 1.5 kb fragment was recovered using a clean-up kit (High Pure PCR-Roche) and ligated into pKSopdE digested with XbaI to obtain pKSopdEGM. The ligation products were transformed into E. coli DH5α competent cells, and the bacteria were plated on LB agar plates containing 15 μg/ml gentamicin. The Enterobacter opdE-strain was transformed with pKSopdEGM to obtain Enterobacter pKSopdEGM; the presence of opdE was confirmed by PCR, and the specific activity of the strain was determined.



Expression and purification

Once the complete sequence of the *opdE* gene was obtained, the primers opdEFW-(5'-TGGGATCCAT GAGCAAAAAACGCAGC-3') (sense, the *BamHI* site is underlined and the start codon is marked in bold) and opdERV-5'-GTCGACTAAGTGGATCTCCGC GCTGAG-3' (antisense, the SalI site is underlined and the termination codon is marked in bold) were designed to amplify the open reading frame using the bacterial genomic DNA as a template. The amplified DNA fragment, opdE, was digested and ligated into the pJET vector and subcloned into the pET-22b (+) vector (Novagen), which encodes a C-terminal hexahistidine tag for affinity purification. The recombinant plasmids were transformed into E. coli BL21 (DE3), and positive transformants were cultured in 100 ml LB medium with 100 μg/ml ampicillin at 30°C with shaking. When the cell density had reached $OD_{600} =$ 0.5, IPTG was added to the culture at a final concentration of 1 mM. After a 4-h induction, the cells were harvested by centrifugation at $12,000 \times g$. The expression level of the protein was assessed by analysis on a 12% SDS-PAGE gel.

The periplasmic protein from E. coli was first extracted by previously described methods (Dresler et al. 2006). The culture pellet was obtained by centrifugation $(10,000 \times g, 5 \text{ min}, 4^{\circ}\text{C})$ and then exposed to osmotic shock by suspension in 1 ml of ice-cold 20% sucrose in 30 mM Tris-HCl, pH 8. Following 20 min incubation at 37°C, the spheroplasts were pelleted $(10,000 \times g, 10 \text{ min}, 4^{\circ}\text{C}),$ resuspended in 1 ml of ice-cold 5 mM MgSO₄ and incubated on ice for 10 min. After centrifugation $(10,000 \times g, 10 \text{ min}, 4^{\circ}\text{C})$, the supernatant (total periplasmic proteins [PP]) was stored at 4°C for subsequent purification. The total PP was passed through 4 ml of Ni-NTA His-Bind Resin (Novagen), which had been equilibrated with 1× Ni-NTA Bind buffer (50 mM NaH₂PO₄, pH 8.0, 300 mM NaCl, and 10 mM imidazole). The column was washed with 1× Ni-NTA Wash buffer (50 mM NaH₂PO₄, pH 8.0, 300 mM NaCl, and 20 mM imidazole) and then the target protein was eluted in four steps by the addition of 0.5 ml 1× Ni-NTA Elution buffer (50 mM NaH₂PO₄, pH 8.0, 300 mM NaCl, and 250 mM imidazole). The eluted fractions were collected separately and analyzed on an SDS-PAGE gel.

Results

Cell growth and degradation of MP in liquid cultures by the Cons002 strain

The cell growth of the Cons002 strain, expressed in viable counts (CFU/ml), with and without MP is shown in Fig. 1. After the first 20 h of incubation, the Cos002 strain grew in the presence of the MP pesticide but not in the absence of the pesticide. This result suggests that Cons002 likely requires a pesticide as an alternative nutrient source.

The degradation kinetics of MP and glucose, as well as the production of PNP, were assessed simultaneously. The ability of Cons002 to degrade MP was measured by HPLC analysis during bacterial growth. Both glucose and MP levels decreased from the initial levels, while PNP, the degradation product of MP, increased proportionally during the first 20 h until the PNP could not be degraded further (Fig. 2). However, the results suggest that other metabolites like dimethylthiophosphoric acid could be use as nutrients sources.

Removal of other OP pesticides in liquid culture

The Cons002 strain removed OPs, MP, parathion and phorate in 28 h. It is interesting to note that Cons002 can significantly degrade MP (71%, 14.2 mg/l), but it only slightly removes parathion (17%, 3.5 mg/l) and phorate (21%, 4.1 mg/l) (Fig. 3). Cadusafos and butiphos could not be removed from liquid medium. These results show that Cons002 probably has the

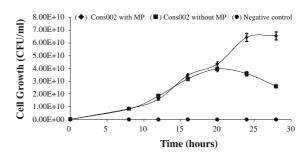


Fig. 1 Cell growth of Cons002 in the presence of MP (20 mg/l) (filled diamond) or in the absence of MP (filled square). Control MM containing 50 mg/l glucose without cells (filled circle). The samples were incubated at 30°C for 28 h as described in the "Materials and methods" section. Each value is the mean of three replicates with *error bars* representing the standard deviation of the mean



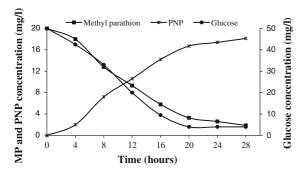


Fig. 2 Loss of MP (filled diamond), loss of glucose (filled circle), and production of PNP (asterisks) throughout the growth (for 28 h at 30°C) of the Cons002 strain cultivated in MM supplemented with 50 mg/l glucose and MP (20 mg/l) at an initial OD_{600} of 0.05

ability to break down the P–S bond of some organophosphate pesticides but not the P–O bond.

Bacterial identification

The Cons002 isolate was circular, smooth and yellow-white. It is a gram-negative bacteria strain. When the partial 16S rRNA gene of Cons002 was compared to previously published sequences in the EMLB database, the sequence displayed the highest identity (99%) with the 16S rRNA gene of an *Enterobacter* sp. (NCBI database, EF471901.1). PFGE was performed to identify the number of replications and the genome size in the *Enterobacter* sp. The genome size was estimated using the genomes of *Mesorhizobium loti*, *Schizosaccharomyces pombe* and *Hansenula wingei* as molecular weight markers for comparison to the

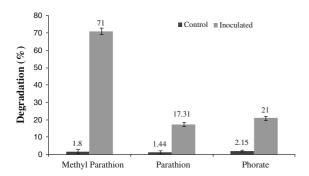


Fig. 3 Biodegradation of OP pesticides by Cons002 in liquid culture. All pesticides were added at an initial concentration of 20 mg/l. The residual concentration was measured by HPLC at 48 h as described in the "Materials and methods" section. Uninoculated media were run as controls. The data are the means \pm standard deviation of three replications

Enterobacter sp. strain Cons002 genome (4.87 Mb) (data not shown).

Construction of a genomic library

The first genes known to be responsible for degrading MP have been isolated from bacterial plasmids, but some bacterial genes, including those from P. monteilii, Plesimonas sp. strain M6, Burkholderia sp. strain NF100, Pseudomonas putida and Stenotrophomonas sp. strain SMSP-1 (Singh and Walker 2006; Horne et al. 2002; Tago et al. 2006; Seo et al. 2007), have been cloned from genomic DNA. The genomic DNA extracted from Cons002 strain was used to generate a library consisting of 5000 clones. Considering a genome size of ~ 4.87 Mb for the Cons002 strain and an average Enterobacter gene size of ~ 1 kb, the theoretical genome coverage of this library is $\sim 3 \times$.

Clones were assayed for MP hydrolysis activity, yielding one positive clone (Hydro1257) that demonstrated the ability to degrade MP (Specific activity = 1.061 ± 0.022 mU/mg protein). The control *E. coli* XL1-Blue strain, with or without pKS+, showed no detectable activity towards MP (Fig. 4). A molecular analysis showed that the Hydro1257 clone contains a fragment of 1156 bp, which was sequenced in both directions.

The OpdE, a novel bacterial Opd-related enzyme

The clone Hydro1257 contains one open reading frame of 759 bp that is 95% identical, at the amino

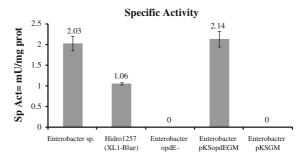


Fig. 4 Specific activity of *Enterobacter* sp. Cons002 (2.03 ± 0.18 mU/mg protein), clone Hydro1257 strain XL1-Blue (1.061 ± 0.022 mU/mg protein) and *Enterobacter* pKSopdEGM (2.14 ± 0.185 mU/mg protein) cultures grown in MM for 48 h at 30°C. Controls: *Escherichia coli* XL1-Blue, *Enterobacter* opdE- and *Enterobacter* pKSGM. Data are the means \pm standard deviations of three replications



acid level, to an endonuclease/exonuclease/phosphatase of *Enterobacter cloacae* subsp. cloacae ATCC 13047 (GenBank Accession no. YP_003613430). The gene is called *opdE* (Organophosphates Degrading from Enterobacter) and encodes a hydrolase of organophosphates (OpdE).

Through sequence comparison with bacterial hydrolase enzymes retrieved from GenBank http://www.ncbi.nlm.nih.gov, we identified the conserved motifs, including the putative COG3568 metal hydrolase motif, which is found in a large number of functionally diverse enzymes, mainly from prokaryotes. The COG3568 motifs is related to pfam03372, a motif found in a family of Mg²⁺-dependent bacterial endonuclease/exonuclease/phosphatase enzymes, and to COG0708 XthA, a motif found in exonuclease III family.

At the amino acid level, the OpdE protein shows slight similarity with proteins reported to degrade organophosphate pesticides, such as OPH protein encoded by opd, opch2 and ophB genes (10, 8 and 4%, respectively); the MPH protein encoded by mpd gene (5%) and the OPAA protein encoded by opaA gene (10%) (Data not show). Furthermore, to clarify the phylogenetic relationship of the OpdE with other enzymes that degrading organophosphates, a neighbor joining phylogenetic tree was constructed using the amino acid sequences of the OPH from B. diminuta and Flavobacterium spp., OP-degrading enzyme (OPDA) from Agrobacterium radiobacter, amidohydrolase superfamily, MPH from the *Pseudomonas* sp. WBC-3, and organophosphorus acid anhydrolase (OPAA) from Alteromonas undina and Alteromonas haloplanktis. As shown in Fig. 5, OpdE formed a distinct group, which is located closest to the branch of putative OPAA (accession number U29240 and U56398) of Alteromonas and also, OPDA (accession number FJ550130) from Lactobacillus brevis. Together, our results on hydrolysis activity, protein domains, and phylogenetic analysis, clearly indicate that OpdE is a new group of enzyme family that degrading organophosphates.

A defective opdE mutant

Enterobacter sp. Cons002 that contains a kanamycinresistance cassette insertion in opdE showed no specific activity towards a MP substrate (Fig. 4), and it was unable to grow with MP as a nutrient source (Fig. 6). This confirms that OpdE is the protein required for *Enterobacter* sp. Cons002 to grow with MP.

Complementation of *Enterobacter* opdE-mutant strain

To confirm the role of the *opdE* gene in MP degradation, the *Enterobacter* opdE-mutant strain was complemented with the construct pKSopdEGM. The transformed strain (called *Enterobacter* pKSopdEGM) was selected with gentamicin. Functionality of the *opdE* gene was confirmed because the complemented mutant strain demonstrated hydrolysis activity. The *Enterobacter* pKSopdEGM strain was tested by measuring the hydrolysis activity and a specific activity = 2.14 ± 0.185 mU/mg protein was obtained (Fig. 4).

Expression of the *opdE* gene in *E. coli* BL21 (DE3)

To display the gene product and to confirm the molecular weight of OpdE, the expression of the opdE gene in $E.\ coli$ BL21 (DE3) cells was analyzed by SDS-PAGE (Fig. 7). The molecular weight of the induced band was ~ 25 kDa, which is consistent with the amino acid sequence of the opdE gene. The specific activity of the crude enzyme extracts of $E.\ coli$ BL21 (DE3) carrying the opdE gene was 8.32 mU/mg protein.

Discussion

Biodegradation by microorganisms is the primary means for the elimination of organophosphate insecticides released to the environment, and several bacterial strains are capable of completely or partially degrading organophosphorus. Some bacteria have been isolated from soils contaminated with organophosphate compound pesticides (Singh and Walker 2006; Kim et al. 2009; Wang et al. 2010). Bacteria capable of degrading MP have been obtained from agricultural soils by the enrichment culture technique (Yáñez-Ocampo et al. 2009). Species identification by 16S rDNA sequence analysis revealed that the isolated strain is related to genus *Enterobacter*. Some species of this genus have been reported to degrade different



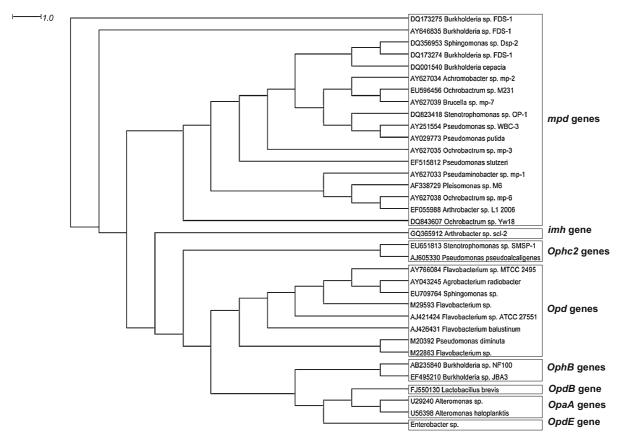


Fig. 5 Phylogenetic reconstruction of the organophosphorus-degrading enzymes family. Phylogenetic tree was constructed from a single multiple alignment using the Phylyp package V

3.57. The *boxes* indicate distinct clusters of organophosphorus-degrading enzymes. For each sequence, NCBI accession number and species are shown

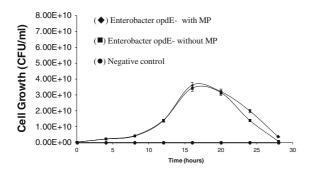


Fig. 6 Cell growth of *Enterobacter* opdE- in the presence of MP (20 mg/l) (*filled diamond*) or the absence of MP (*filled square*). MM containing 50 mg/l glucose without cells was used as a control (*filled circle*). The samples were incubated at 30°C for 28 h as described in the "Materials and methods" section. Each value is the mean of three replicates, and error bars represent the standard deviation of the mean

contaminants, such as chlorpyrifos (Singh et al. 2004); 2,4,6-trinitrotoluene (Robertson and Jjemba 2005); di- and trichlorobenzenes (Adebusoyea et al. 2007);

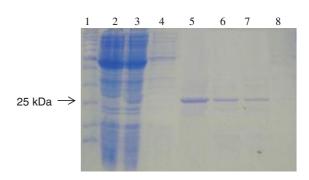


Fig. 7 SDS-PAGE gel electrophoresis of the purified protein. *Lane 1* protein marker (Fermentas), *lane 2 E. coli* transformed with the pET-22b (+) plasmid, *lane 3* PP with expressed OpdE, *lane 4* Ni-NTA wash, *lanes 5*, *6*, *7* and *8* Ni-NTA elutions

products of the highly toxic organophosphate nerve agents sarin, soman, GF, VX and rVX (Ghanem et al. 2007) and petroleum (Kaoa et al. 2010).



Enzymes that can hydrolyze OP compounds have been identified and characterized after isolation from different microbial species (Singh and Walker 2006). The *opdE* gene is different from genes encoding organophosphorus hydrolases deposited in GenBank, which indicates that the enzyme is a novel organophosphorus hydrolase. The use of enzymes for the detection and decontamination of OPs and for medical applications is of great interest (Sogorb and Vilanova 2002; Singh 2009; Theriot and Grunden 2010; Van and Pletschke 2010).

No natural substrates have been identified to OPs hydrolases (Afriat et al. 2006), and by the other side, there is a rich literature addressing the evolutionary potential by promiscuous activities (Seibert and Raushel 2005). It is widely accepted that many enzymes evolved via gene duplication (Hurles 2004; Thoden et al. 2004) and that some of they are capable of catalyzing other reactions, in addition to the ones for which they are physiologically specialized. Additionally, some publications showed that mutations increase the efficiency of a promiscuous activity by changes in the catalytic site or genetic expression. This fact allows to bacteria some physiologically and evolutionary advantages (O'Brien and Herschlag 1999; DePristo 2007; Khersonsky and Tawfik 2010). By the other hand, several groups have suggested that OP hydrolases diverged from one pre-existing progenitor protein (DePristo 2007; Khersonsky and Tawfik 2010).

The OPH enzymes encoded by the *opd* and *mpd* genes are members of the amidohydrolase superfamily, a group of enzymes sharing the same $(\beta/\alpha)_8$ -barrel structural fold that catalyzes the hydrolysis of amide or ester functional groups at carbon and phosphorus centers (Seibert and Raushel 2005). These centers might have evolved from PTE-like lactonase (PLL) enzymes, which hydrolyze lactones and have low activity towards OPs (Afriat et al. 2006). Another possibility is that they evolved from the phosphotriesterase homology protein (PHP) of *E. coli*, which is a metalloenzyme with a $(\beta/\alpha)_8$ -barrel structure (Scanlan and Reid 1995).

Because these families of proteins are different (Fig. 5) in regard to scaffolding, catalytic mechanisms and the chemical structures of target lactones and OPs, it was proposed that a promiscuous phosphotriesterase of lactonases emerged from the overlap between the transition states of lactones and OPs during hydrolysis

(Afriat et al. 2006). Other hydrolases with similar transition states, such as esterases or prolidases (like OPAA), also have promiscuous phosphotriesterase activity, supporting this hypothesis (Elias et al. 2008).

Several studies of proteins reveal families of enzymes; non-homologous enzymes evolve in separate biological contexts to catalyse the same or similar biochemical transformation. Often such enzymes have nothing in common beyond their function (Gherardini et al. 2007, Pratt and McLeish 2010). As shown in the phylogenetic tree, there are several groups of proteins that have, probably, different ancestors, but can degrade OP compounds. OpdE protein has endonuclease and exonuclease domains and too degrade OPs, suggesting that evolution is according to their function. These studies are the beginning to find many more progenitors for others degrading OPs.

Some bacteria, such as *Flavobacterium* sp. ATCC 27551 (diazinon), *Burkholderia* sp. NF100 (fenitrothion), *Arthrobacter* sp. scl-2 (isocarbophos), and *Enterobacter* sp. strain B14 (chlorpyrifos) (Sethunathan and Yoshida 1973; Tago et al. 2006; Rong et al. 2009; Singh et al. 2004) have been reported to utilize OPs as a nutrients source. However, when *Enterobacter* sp. Cons002 was cultured in the presence of MP without glucose, the bacteria did not show any growth (data not shown).

Data of this work show that *Enterobacter* sp. Cons002 uses MP as a nutrient source, probably the dimethyl thiophosphoric acid, as growth was observed despite the depletion of glucose. In contrast, *Enterobacter* opdE-could not degrade MP, and therefore growth was affected when glucose in the culture medium was depleted (Fig. 6). Additionally, *Enterobacter* sp. Cons002 was not able to grow after the depletion of glucose without MP (Fig. 1). These results suggest that MP is degraded in a cometabolism process.

The *opdE* gene from Cons002 was cloned and expressed in *E. coli*, and this is the first report of the recombinant expression of a gene from the genus *Enterobacter* that encodes a protein that degrades OPs.

In conclusion, we isolated *Enterobacter* sp. Cons002, which is able to degrade the pesticides MP, parathion and phorate and probably uses these pesticides as a nutrient source. We characterized the *opdE* gene, which encodes a hydrolase enzyme with the capacity to hydrolyze MP. The phylogenetic analysis suggests that the organophosphorus-degrading functions of OpdE might have evolved from



different progenitors than those previously reported. This kind of enzyme is thought to play an important role in the degradation of organophosphate compounds and has many promising applications, including the bioremediation of materials polluted by OPs, the development of biosensors for OPs and therapies for OP poisoning.

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References

- Adebusoyea SA, Picardalb FW, Iloria MO, Amunda OO, Fuquac C, Greindlec N (2007) Aerobic degradation of di- and trichlorobenzenes by two bacteria isolated from polluted tropical soils. Chemosphere 66:1939–1946. doi:10.1016/j.chemosphere.2006.07.074
- Afriat L, Roodveldt C, Manco G, Tawfik DS (2006) The latent promiscuity of newly identified microbial lactonases is linked to a recently diverged phosphotriesterase. Biochemistry 45:13675–13686. doi:10.1021/bi061268r
- Alexeyev MF, Shokolenko IN, Croughan TP (1995) Improved antibiotic-resistance gene cassettes and omega elements for *Escherichia coli* vector construction and in vitro deletion/insertion mutagenesis. Gene 160:63–67. doi:10.1016/ 0378-1119(95)00108-I
- Brown K (1980) The phosphotriesterase of *Flavobacterium* sp. Soil Biol Biochem 12:105–112. doi:10.1016/0038.0717 (80)90044-9
- Chen W, Richins R, Mulchandani P, Kaneva I, Mulchchandani A (2000) Biodegradation of organophosphorus nerve agents by surface expressed organophosphorus hydrolase. In: Zwanenburg B, Mikalajezyk M, Kielbasiski P (eds) Enzymes in action green solution for chemical problems, vol 33. Kluwer Academic, Dordrecht, pp 211–221
- Cui Z, Li S, Fu G (2001) Isolation of methyl-parathiondegrading strain M6 and cloning of the methyl-parathion hydrolase gene. Appl Environ Microbiol 67:4922–4925. doi:10.1128/AEM.67.10.4922-4925.2001
- Datsenko KA, Wanner BL (2000) One-step inactivation of chromosomal genes in *Escherichia coli* K-12 using PCR products. Proc Natl Acad Sci 97:6640–6645. doi:10.1073/pnas.120163297
- DePristo MA (2007) The subtle benefits of being promiscuous: adaptive evolution potentiated by enzyme promiscuity. HFSP J 1:94–98. doi:10.2976/1.2754665
- Dresler K, Van DHJ, Müller RJ, Deckwer WD (2006) Production of a recombinant polyester-cleaving hydrolase from Thermobifida fusca in *Escherichia* coli. Bioprocess Biosyst Eng 29:169–183. doi:10.1007/s00449-006-0069-9
- Elias M, Dupuy J, Merone L, Mandrich L, Porzio E, Moniot S, Rochu D, Lecomte D, Rossi M, Masson P, Manco G, Chabriere E (2008) Structural basis for natural lactonase

- and promiscuous phosphotriesterase activities. J Mol Biol 5:1017–1028. doi:10.1016/j.jmb.2008.04.022
- Ghanem E, Raushel FM (2005) Detoxification of organophosphate nerve agents by bacterial phosphotriesterase. Toxicol Appl Pharmacol 207:459–470. doi:10.1016/j.taap.2005.02.025
- Ghanem E, Li Y, Xu C, Raushel FM (2007) Characterization of a phosphodiesterase capable of hydrolyzing EA 2192, the most toxic degradation product of the nerve agent VX. Biochemistry 46:9032–9040. doi:10.1021/bi700561k
- Gherardini PF, Wass MN, Helmer-Citterich M, Sternberg MJE (2007) Convergent evolution of enzyme active sites is not a rare phenomenon. J Mol Biol 372:817–845. doi:10.1234/12345678
- Horne I, Harcourt RL, Sutherland TD, Russell RJ, Oakeshott JG (2002) Isolation of a *Pseudomonas monteilli* strain with a novel phosphotriesterase. FEMS Microbiol Lett 206:51–55. doi:10.1111/j.1574-6968.2002.tb10985.x
- Hurles M (2004) Gene duplication: the genomic trade in spare parts. PLoS Biol. 2(7): e206. doi:10.1371/journal.pbio. 0020206
- Kaoa CM, Chenb CS, Tsab FY, Yangc KH, Chiend CC, Liangb SH, Yangb CA, Chenb SC (2010) Application of real-time PCR, DGGE fingerprinting, and culture-based method to evaluate the effectiveness of intrinsic bioremediation on the control of petroleum-hydrocarbon plume. J Hazard Mater 178:409–416. doi:10.1016/j.jhazmat.2010.01.096
- Khersonsky O, Tawfik DS (2010) Enzyme promiscuity: a mechanistic and evolutionary perspective. Annu Rev Biochem 79:11.1–11.35. doi:10.1146/annurev-biochem-030409-143718
- Kim KD, Ahn JH, Kim T, Park SC, Seong CN, Song HG, Ka JO (2009) Genetic and phenotypic diversity of fenitrothiondegrading bacteria isolated from soils. J Microbiol Biotechnol 19:113–120. doi:10.4014/jmb.0808.467
- Lopez I, Ruíz LF, Cocolin L, Orr E, Phister T, Marshall M, VanderGheynst J, Mills DA (2003) Design and Evaluation of PCR Primers for Analysis of Bacterial Populations in Wine by Denaturing Gradient Gel Electrophoresis. Appl Environ Microbiol 69:6801–6807. doi:10.1128/AEM.69. 11.6801-6807.2003
- Lowry OH, Rosebrough NJ, Farr AL, Randall RJ (1951) Protein measurement with the folin phenol reagent. J Biol Chem 193:265–275
- Miller SA, Dykes DD, Polesky HF (1988) A simple salting out procedure for extracting DNA from human nucleated cells. Nucleic Acids Res 16:1215. doi:10.1093/nar/16.3.1215
- O'Brien PJ, Herschlag D (1999) Catalytic promiscuity and the evolution of new enzymatic activities. Chem Biol 6(4): R91–R105. doi:10.1016/S1074-5521(99)80033-7
- Ortiz-Hernández ML, Quintero-Ramírez R, Nava-Ocampo AA, Bello-Ramírez AM (2003) Study of the mechanism of *Flavobacterium* sp for hydrolyzing organophosphate pesticides. Fundam Clin Pharmacol 17:717–722. doi:10.1046/j.1472-8206.2003.00202.x
- Porzio E, Merone L, Mandricha L, Rossia M, Manco G (2007) A new phosphotriesterase from *Sulfolobus acidocaldarius* and its comparison with the homologue from *Sulfolobus solfataricus*. Biochemie 89:625–636. doi:10.1016/j.biochi. 2007.01.007



- Pratt RF, McLeish MJ (2010) Structural relationship between the active sites of β -lactam-recognizing and amidase signature enzymes: convergent evolution? Biochemistry 49(45):9688–9697. doi:10.1021/bi1012222
- Robertson BK, Jjemba PK (2005) Enhanced bioavailability of sorbed 2,4,6-trinitrotoluene (TNT) by a bacterial consortium. Chemosphere 58:263–270. doi:10.1016/j.chemosphere.2004.
- Rong L, Guo X, Chen K, Zhu J, Li S, Jiang J (2009) Isolation of an isocarbophos-degrading strain of *Arthrobacter* sp. scl-2 and identification of the degradation pathway. J Microbiol Biotechnol 19:1439–1446. doi:10.4014/jmb.0811.0626
- Sambrook J, Russell DW (2001) Molecular cloning. A laboratory manual. Cold Spring Harbor, New York
- Scanlan TS, Reid RC (1995) Evolution in action. Chem Biol 2:71–75. doi:10.1016/1074-5521(95)90278-3
- Seibert CM, Raushel FM (2005) Structural and catalytic diversity within the amidohydrolase superfamily. Biochemistry 44:6383–6391. doi:10.1021/bi047326v
- Seo JS, Keum YS, Harada RM, Li QX (2007) Isolation and characterization of bacteria capable of degrading polycyclic aromatic hydrocarbons (PAHs) and organophosphorus pesticides from PAH-contaminated soil in Hilo, Hawaii. J Agric Food Chem 55:5383–5389. doi:10.1021/jf0637630
- Sethunathan N, Yoshida T (1973) A *Flavobacterium* sp that degrades diazinon and parathion. Can J Microbiol 19:873–875. doi:10.1139/m73-138
- Shen YJ, Lu P, Mei H, Yu HJ, Hong Q, Li SP (2010) Isolation of a methyl-parathion-degrading strain *Stenotrophomonas* sp. SMSP-1 and cloning of the *ophc2* gene. Biodegradation 21:785–792. doi:10.1007/s10532-010-9343-2
- Singh B (2009) Organophosphorus-degrading bacteria: ecology and industrial applications. Nat Rev Microbiol 7:156–164. doi:10.1038/nrmicro2050
- Singh B, Walker A (2006) Microbial degradation of organophosphorus compounds. FEMS Microbiol Rev 30(3): 428–471. doi:10.1111/j.1574-6976.2006.00018.x
- Singh B, Walker A, Morgan J, Wright D (2004) Biodegradation of chlorpyrifos by *Enterobacter* Strain B-14 and its use in bioremediation of contaminated soils. Appl Environ

- Microbiol 70:4855–4863. doi:10.1128/AEM.70.8.4855-4863
- Smith CL, Cantor CR (1987) Purification, specific fragmentation, and separation of large DNA molecules. Methods Enzymol 155:449–467. doi:10.1016/0076-6879(87)55050
- Sogorb MA, Vilanova E (2002) Enzymes involved in the detoxification of organophosphorus, carbamate and pyrethroid insecticides through hydrolysis. Toxicol Lett 128(1–3):215–228. doi:10.1016/S0378-4274(01)00543-4
- Tago K, Yonezawab S, Ohkouchib T, Hashimotoc M, Hayatsub M (2006) Purification and characterization of fenitrothion hydrolase from *Burkholderia* sp. NF100. J Biosci Bioeng 101:80–82. doi:10.1263/jbb.101.80
- Theriot CM, Grunden AM (2010) Hydrolysis of organophosphorus compounds by microbial enzymes. Appl Microbiol Biotechnol 89:35–43. doi:10.1007/s00253-010-2807-9
- Thoden JB, Ringia EAT, Garrett JB, Gerlt JA, Holden HM, Rayment I (2004) Evolution of enzymatic activity in the enolase superfamily: structural studies of the promiscuous *o*-succinylbenzoate synthase from *Amycolatopsis*. Biochem 43:5716–5727. doi:10.1021/bi0497897
- Van DJS, Pletschke B (2010) Review of the use of enzymes for the detection of organochlorine, organophosphate and carbamate pesticides in the environment. Chemosphere 82:291–307. doi:10.1016/j.chemosphere.2010.10.033
- Wang L, Wen Y, Guo X, Wang G, Li S, Jiang J (2010) Degradation of methamidophos by *Hyphomicrobium* species MAP-1 and the biochemical degradation pathway. Biodegradation 21:785–792. doi:10.1007/s10532-009-9320-9
- Yáñez-Ocampo G, Sanchez-Salinas E, Jimenez-Tobon GA, Penninckx M, Ortiz-Hernández ML (2009) Removal of two organophosphate pesticides by a bacterial consortium immobilized in alginate or tezontle. J Hazard Mater 168:1554–1561. doi:10.1016/j.hazmat.2009.03.047
- Zhang R, Cui Z, Zhang X, Jiang J, Gu JD, Li S (2006) Cloning of the organophosphorus pesticide hydrolase gene clusters of seven degradative bacteria isolated from a methyl-parathion contaminated site and evidence of their horizontal gene transfer. Biodegradation 5:465–472. doi:10.1007/s10532-005-9018-6

